

Exceptional Perrhenate/Pertchnetate Uptake and Subsequent Immobilization by a Low-Dimensional Cationic Coordination Polymer: Overcoming the Hofmeister Bias Selectivity

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1 **Exceptional Perrhenate/Pertechnetate Uptake and Subsequent**
2 **Immobilization by a Low-Dimensional Cationic Coordination**
3 **Polymer: Overcoming the Hofmeister Bias Selectivity**

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16

17 **ABSTRACT**

18 We report one of the most efficient scavenger materials, a cationic crystalline
19 coordination polymer SBN for trapping ReO_4^- , a surrogate for $^{99}\text{TcO}_4^-$ as an anionic
20 radioactive contaminant of great concern. The uptake capacity for ReO_4^- reaches 786
21 mg/g, a value noticeably higher than the state of art anion-exchange resins and other
22 inorganic or hybrid anion sorbents. Once being captured, ReO_4^- is greatly
23 immobilized as almost no ReO_4^- can be eluted using large excess of nitrate, carbonate,
24 and phosphate anions. The processes are featured by a complete and irreversible
25 single-crystal-to-single-crystal structural transformation from SBN to the ReO_4^-
26 incorporated phase (SBR). The coordination environments of NO_3^- and ReO_4^- probed
27 by single crystal structures clearly unravel the underlying mechanism, where each
28 ReO_4^- in SBR binds to multiple Ag^+ sites forming strong Ag–O–Re bonds, and to
29 4,4'-bipyridine through a dense hydrogen bond network. These structural insights lead
30 to a significant difference of solubility product constants between SBN and SBR,
31 which is further confirmed by first principle calculations showing a large binding
32 energy difference of 35.61 kcal/mol. To the best of our knowledge, SBR is the least
33 soluble perrhenate/pertechnetate salt reported, which may be considered as a potential
34 waste form for direct immobilization of TcO_4^- .

35

36 ■ INTRODUCTION

37 ^{99}Tc is a long-lived fission product of ^{235}U or ^{239}Pu and primarily exists in aqueous
38 solution as highly soluble, mobile, and stable TcO_4^- anion. Release of ^{99}Tc is a great
39 concern in used fuel reprocessing, waste management, weapon testing, and nuclear
40 accidents.^{1,2} Technetium contamination has already been identified at several sites
41 such as Hanford in southeastern Washington and Savannah River site in South
42 Carolina.^{3,4} Control of technetium release by waste form design and remediation of
43 technetium pollution in the aqueous environment are therefore highly desirable, but
44 are in fact of great challenge originating from the non-complexing nature of TcO_4^-
45 anion.⁵ Currently in Nuclear Waste Treatment and Immobilization, vitrification is the
46 primary approach.⁶ Because of the high temperature of vitrification, ^{99}Tc is volatilized
47 and therefore only a portion is captured in the vitrified LAW waste form.⁷ One
48 proposed strategy is the chemical reduction from Tc(VII)O_4^- to Tc(IV) and the
49 separation from the aqueous solution is achieved by formation of insoluble Tc(IV)
50 complexes.⁸⁻¹¹ These can be further incorporated into a pre-designed waste form
51 containing crystallographically compatible sites for accommodating Tc(IV) .¹²⁻¹⁴ The
52 most significant issue for this strategy is the re-oxidation of Tc(IV) back to Tc(VII)
53 followed by the subsequent escaping from the host material.^{14,15} Although many
54 attempts have been focused on the kinetic retention of Tc(IV) in the lattice of waste
55 form,^{12,14,16} the re-oxidation seems to be inevitable during the long geological disposal
56 period given this process is thermodynamically favorable under aerobic conditions.

57 This issue calls for the possibility of technetium disposal directly in the form of
58 Tc(VII)O_4^- . Although this has been thought to be not applicable in light that almost all
59 fifty known pertechnetate salts are sparingly soluble,¹⁷ a recent study shows that
60 pertechnetate can be incorporated into the lattice of sodalite as a potential waste
61 form.¹⁸⁻²¹ However, because of smaller ionic size of nitrate, nitrate anion possesses a
62 clear preference to enter the sodalite cage. In addition, although a variety of
63 anion-exchange materials have already been investigated for removing TcO_4^- from
64 either nuclear waste stream or contaminated aqueous solutions including the state of

65 art anion-exchange polymeric organic resins,²²⁻²⁴ molecular and supramolecular
66 receptors,²⁵ layered double hydroxides (LDHs),^{26,27} metal-organic frameworks
67 (MOFs),^{28,29} and porous organic materials,³⁰ they cannot be used for further
68 immobilization of TcO_4^- , given these anion-exchange reactions are typically
69 reversible.

70 Recently, Oliver et al³¹ reported a low-dimensional cationic coordination polymer
71 $[\text{Ag}(\text{bipy})]\text{NO}_3$ (denoted as SBN, bipy= 4,4'-bipyridine) used as an effective
72 adsorbent for the removal of ClO_4^- from aqueous solution with a very high capacity.
73 In this work, we document our investigations of this type of material containing
74 abundant coordination-available open Ag^+ site that can selectively bind to
75 $\text{ReO}_4^-/\text{TcO}_4^-$, leading to capabilities of not only removing $\text{ReO}_4^-/\text{TcO}_4^-$ from aqueous
76 solution in a highly effective manner, but also greatly immobilizing $\text{ReO}_4^-/\text{TcO}_4^-$ in its
77 crystal lattice even in the presence of NO_3^- in large excess. ReO_4^- is used as a
78 chemical surrogate for TcO_4^- in this study given their almost identical charge densities
79 and many thermodynamic parameters except for the redox behaviors (**Table S1**).³²
80 The trapping mechanism was elucidated by a combination of single crystal X-ray
81 diffraction analysis, thermodynamic analysis, and density function theory (DFT)
82 analysis on binding. This work represents an in situ recrystallization route for TcO_4^-
83 removal from aqueous by low-dimensional cationic MOFs and demonstrates the
84 promising for this type of materials for technetium disposal.

85

86 ■ MATERIALS AND METHODS

87 **Synthesis and Characterization.** SBN material was synthesized by hydrothermal
88 method, as reported before (details were provided in the Supporting Information).^{31,33}
89 The detailed synthesis routes of other cationic materials for comparison, including
90 NDTB-1,³⁴ Mg-Al-LDH,³⁵ and $\text{Yb}_3\text{O}(\text{OH})_6\text{Cl}\cdot 2\text{H}_2\text{O}$ ³⁶ were described in the
91 Supporting Information. The commercial Purolite A532E anion-exchange resin was
92 obtained from Purolite Co., Ltd. and dried at 60 °C for 24 h before use.

93 **Batch Experiments.** All the experiments were conducted at 25 °C using the batch

94 sorption method. The solid/liquid ratio performed in all batch experiments was 0.5
95 g/L. In a typical experiment, 10 mg of SBN was added into 20 mL of aqueous
96 solution containing certain amounts of ReO_4^- . The resulting mixture was stirred for a
97 desired contact time and separated with a 0.22 μm nylon membrane filter. The
98 concentrations of ReO_4^- in aqueous solution were determined by Inductively Coupled
99 Plasma Mass Spectrometry (ICP-MS, Thermo Scientific) and/or Inductively Coupled
100 Plasma Optical Emission Spectrometry (ICP-OES, Thermo Fisher Scientific iCAP
101 7000). Single crystals of SBR were obtained directly from batch experiment when
102 ReO_4^- was in small excess. After anion-exchange, the compounds were washed with
103 deionized water several times, air-dried, and then characterized by Fourier Transform
104 Infrared Spectroscopy (FT-IR) spectroscopy, Single-crystal X-ray Diffraction
105 (SCXRD), Powder X-ray Diffraction (PXRD), and Scanning Transmission Electron
106 Microscopy combined with an Energy Dispersive X-ray Spectrometer (STEM-EDS,
107 Tecnai G2 spirit BioTwin).

108 **SBN and SBR Dissolution Experiments.** To determine the solubility product
109 constant of the samples, the dissolution experiments were measured. Typically, 10 mg
110 of sample (SBN or SBR) was added into 20 mL of aqueous solutions at various pHs.
111 The concentrations of Ag^+ in aqueous solution were determined by ICP-MS. The
112 thermodynamic parameters were calculated based on the solubility data. More details
113 about the analyses are provided in the Supporting Information.

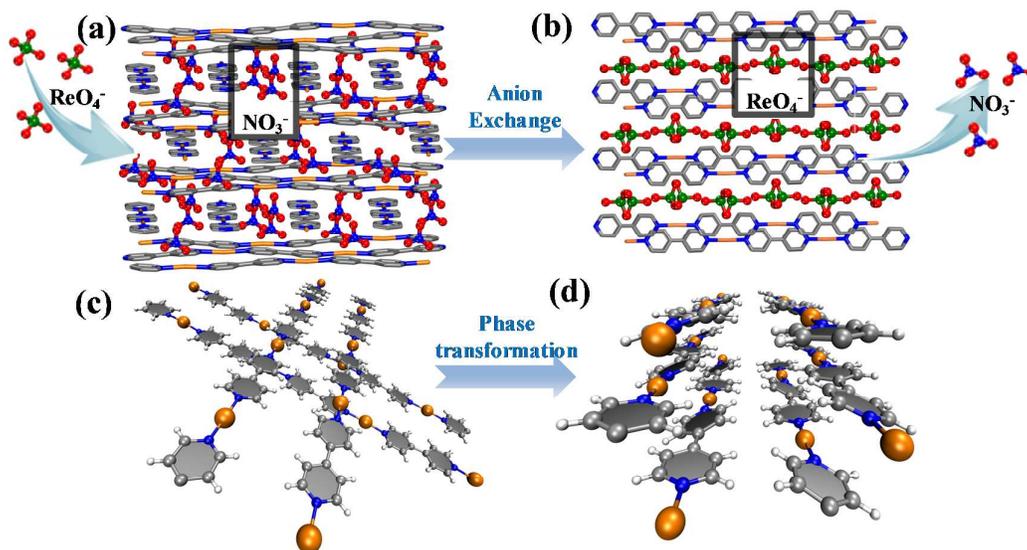
114 **Computational Method.** First-principle calculations based on density functional
115 theory (DFT) were performed using Gaussian 09 program.³⁷ Two fragments of SBN
116 and SBR extracted from the experimentally obtained crystal structures were chosen as
117 the computational models. Geometry optimizations were performed at
118 M062X-D3/SDD~6-31G(d, p) level, in which the M062X³⁸ functional combining
119 with the D3 version of Grimme's dispersion³⁹ were employed, and the
120 Stuttgart/Dresden relativistic effective core potentials (ECPs) with corresponding
121 valence basis set (SDD)⁴⁰ and the standard Gaussian-type 6-31G(d, p) basis set⁴¹ were
122 applied for heavy atoms (Ag and Re) and light atoms (C, H, O and N), respectively.

123 The C, N, and Ag atoms of SBN and SBR frameworks were freeze during
124 optimizations. Single-point energies were calculated for the complexes, the
125 frameworks and anions after geometry optimizations at the same level and the binding
126 energies (E_b) between anions and the frameworks were calculated by $E_b = E(\text{complex})$
127 $- E(\text{framework}) - E(\text{anion})$. The basis set superposition error (BSSE) has been
128 excluded in calculating E_b .

129

130 ■ RESULTS AND DISCUSSION

131 **Synthesis and Characterization of SBN.** The single-crystalline material of SBN
132 crystals was synthesized by hydrothermal method.^{31,33} Single-crystal X-ray
133 Diffraction reveals the overall structure of SBN can be best described as a series of
134 one-dimensional $[\text{Ag}(\text{bipy})]^+$ (bipy= 4,4'-bipyridine) chains stacking together along b
135 axis in a crossed manner (**Figures 1, S1a, and b**). Each Ag^+ ion is coordinated by two
136 neutral 4,4'-bipyridine ligands while each 4,4'-bipyridine ligand bridges two Ag^+ ion
137 within the chains with the Ag-N bonds of 2.163 Å. Weakly coordinated NO_3^- reside in
138 the open space as the charge-balancing anion. The hydrolytic stability measurements
139 was conducted before the sorption experiments, showing The SBN crystals are stable
140 within pH range from 3 to 10 (**Figure S2**) while SBN crystals are completely
141 dissolved below pH 2.



142

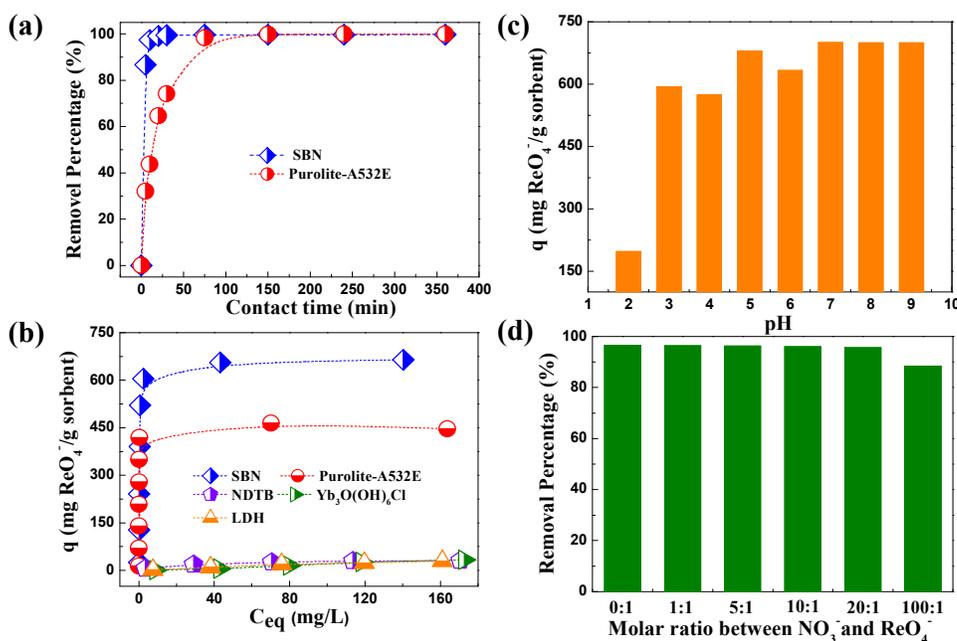
143 **Figure 1.** Crystal structures of SBN (a) and SBR (b) showing a single-crystal-to-single-crystal
144 structural transformation mechanism; The black square regions represent locations of NO_3^- (a) and
145 ReO_4^- (b); Packing modes of $[\text{Ag}(\text{bipy})]^+$ chains of (c) SBN and (d) SBR in the 3D space.

146 **ReO_4^- Sorption experiments.** In order to check the anion-exchange properties of
147 SBN towards TcO_4^- , the sorption experiments were investigated using the chemical
148 surrogate ReO_4^- . Sorption kinetics of ReO_4^- on SBN was investigated as a function of
149 contact time at pH 7 with an initial ReO_4^- concentration of 28 ppm and the solid/liquid
150 ratio of 0.5 mg/ml. As shown in **Figure 2(a)**, the uptake kinetics of ReO_4^- by SBN is
151 extremely fast as the sorption equilibrium can be reached within 10 minutes, where all
152 ReO_4^- was completely removed from aqueous solution. The sorption rate is
153 significantly faster than that of the commercial anion-exchange resin (A532E),
154 specially designed for the removal of ClO_4^- and TcO_4^- under the identical condition,
155 where the sorption equilibrium is reached after at least 2 h.⁴² Compared with other
156 reported MOFs with anion-exchange properties, SBN also possesses clear advantages.
157 For example, the exchange equilibrium of ReO_4^- by UiO-66-NH_3^+ is reached after at
158 least 24 h.⁴³ In practical application of TcO_4^- removal, the ultrafast removal kinetics
159 would significantly decrease the contact time between the sorbents and radioactive
160 waste solution, lowering the extent of material damage induced by radiation and
161 hydrolysis.

162 The uptake isotherm experiments were conducted for 12 h and the results are
163 shown in **Figure 2(b)**. Four other well-known anion sorbent materials, Mg-Al-LDH,
164 NDTB-1, $\text{Yb}_3\text{O}(\text{OH})_6\text{Cl}$, and Purolite A532E were also synthesized to evaluate their
165 sorption properties towards ReO_4^- for comparison. Impressively, the sorption capacity
166 of ReO_4^- by SBN reaches 714 mg ReO_4^-/g sorbent obtained from the fitting results
167 based on the Langmuir model (**Figure S3 and Table S2**). This value is noticeably
168 higher than most of the reported sorbents to date (**Table S3**) including all inorganic
169 cationic materials and organic polymeric based ion exchange resins. When SBN was
170 soaked in a solution with small molar excess of ReO_4^- , the sorption capacity was
171 measured to be 786 mg/g and is close enough to the theoretic value of 767 mg/g,

172 suggesting all NO_3^- in the original material can be fully exchanged. Comparing with
173 few examples of other reported cationic MOFs tested for $\text{ReO}_4^-/\text{TcO}_4^-$ removal, the
174 sorption capacity of SBN is also higher than that of UiO-66- NH_3^+ (159 mg/g),⁴³
175 SLUG-21 (602 mg/g),²⁸ and our recently reported SCU-100 (541 mg/g).⁴⁴

176 We investigated the effect of pH ranging from 2 to 9 on the ReO_4^- sorption onto
177 SBN. As shown in **Figure 2(c)**, the removal capacity increased from pH 2 to 5 and
178 remained almost unchanged afterwards. Although SBN material fully dissolves at pH
179 2, the removal capacity is still quite high at ca. 200 mg/g, initially indicating that the
180 recrystallization based phase transformation mechanism plays a role. The effect of
181 competing NO_3^- amount on the ReO_4^- uptake was also studied given large quantities
182 of NO_3^- is present in the real waste solution or natural water. For example, the molar
183 ratio of $\text{NO}_3^- / \text{TcO}_4^-$ is close to 300 in Hanford low-level waste melter off-gas
184 scrubber stream.²⁶ As shown in **Figure 2(d)**, increasing the amounts of NO_3^- did not
185 result in an obvious decrease of ReO_4^- removal percentage. Even in presence of
186 100-fold NO_3^- , the removal percentage was still as high as 90%, suggesting an
187 excellent exchange selectivity towards ReO_4^- over NO_3^- possessed by SBN, which
188 represents the opposite case of the incorporation of $\text{ReO}_4^-/\text{NO}_3^-$ into sodalite.¹⁸
189 Furthermore, the removal selectivity between ReO_4^- , CO_3^{2-} , H_2PO_4^- , SO_4^{2-} , Cl^- and
190 ClO_4^- was also investigated and the results illustrate that SBN can completely remove
191 ReO_4^- in the presence of other anions (**Figure S4**).



192

193 **Figure 2.** (a) Sorption kinetics data of ReO_4^- into SBN and A532E; (b) Sorption isotherms of
 194 ReO_4^- by SBN compared with LDH, NDTB-1, Purolite-A532E, and $\text{Yb}_3\text{O}(\text{OH})_6\text{Cl}$ materials; (c)
 195 Effect of pH on the sorption of ReO_4^- into SBN, (d) Effect of competing NO_3^- on the uptake of
 196 ReO_4^- by SBN.

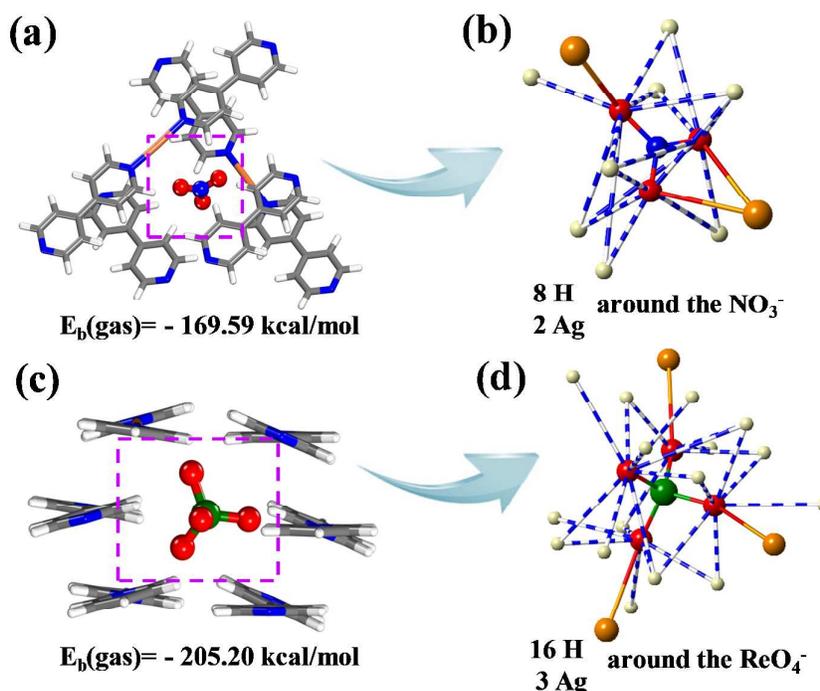
197 **Mechanisms and Thermodynamic study.** Optical microscope photograph and SEM
 198 imaging show that gray block crystals of SBN disintegrates into smaller colorless
 199 needle crystals of SBR (**Figures S5 and S6**), indicating a complete phase transition
 200 occurs after the anion-exchange when ReO_4^- is in small excess. This is further
 201 confirmed by the PXRD measurements (**Figure S7b and S8**) and STEM-EDS
 202 mapping profiles (**Figure S9**) on the samples before and after the anion-exchange. In
 203 FT-IR spectra (**Figure S7a**), the arise of new peaks at 896 cm^{-1} and 860 cm^{-1} as well
 204 as the decrease of the peak at 1326 cm^{-1} is consistent with the complete substitution of
 205 NO_3^- by ReO_4^- in the solid.

206 More importantly, the crystal structure of needle phase SBR can be still clearly
 207 resolved by single-crystal X-ray diffraction (**Table S4**) and significantly deviates from
 208 that of SBN. This is surprising given the single-crystallinity of ion-exchange material
 209 is often ruined during the ion-exchange process. The mechanism of single-crystal

210 (SBN) to single-crystal (SBR) phase transformation is well described in **Figure 1**.
211 although the major framework can be still best described by packing of
212 one-dimensional $[\text{Ag}(\text{bipy})]^+$ chains in 3D space, the packing mode undergoes a
213 transition from the crossed packing in SBN (**Figures 1c, S1a and S1b**) to the parallel
214 packing in SBR (**Figures 1d, S1c and S1d**). By a close examination on the
215 coordination environments of NO_3^- in SBN and ReO_4^- in SBR, we propose that this
216 dramatic change on the overall framework structure is indeed induced by change of
217 charge-balancing anions and in fact occurs to better match the coordination
218 requirement of ReO_4^- . In the structure of SBN (**Figure 3b and Table S5**), the three
219 oxygen atoms within each NO_3^- binds to two Ag^+ cations with the Ag-O bonds
220 ranging from 2.63 to 2.88 Å, and eight H atoms from 4,4'-bipyridine group forming a
221 hydrogen bond network (O-H distances are from 2.21 to 3.93 Å). In comparison, the
222 four oxygen atoms of each ReO_4^- in the structure of SBR (**Figure 3d**) is coordinated
223 by obviously increased amounts (three) of Ag^+ at distances from 2.87 to 3.01 Å and a
224 much denser hydrogen bond network containing 16 H atoms around each ReO_4^- at
225 distances from 2.31 to 3.85 Å. This would result in a deep energy well for the
226 effective stabilization of the ReO_4^- in the structure.

227 To quantitatively interpret the energetic difference between the coordination of
228 NO_3^- and ReO_4^- in these two structures, we performed DFT calculations on the
229 binding energy based on an optimized crystal structure model.³⁷⁻⁴¹ As more H and Ag^+
230 provide greater attractions for the central anions, the calculated binding energy
231 between the anion and the cationic framework in SBR (-205.20 kcal/mol) becomes
232 significantly larger in comparison to that in SBN (-169.59 kcal/mol). This large
233 energetic difference of 35.61 kcal/mol is the intrinsic driving force accounting for
234 large exchange capacity, rapid removal kinetics, and excellent sorption selectivity
235 towards ReO_4^- . Furthermore, in order to qualitatively distinguish and assess the
236 binding contributions from hydrogen bonds and $\text{O}\cdots\text{Ag}^+$ bonds respectively, we also
237 calculated the $\text{O}\cdots\text{H}$ binding energies between the central anions (NO_3^- and ReO_4^-)
238 and the neutral frameworks (without Ag^+) by simply removing the Ag^+ cations in both

239 SBN and SBR fragments (for **Figure 3a** and **3c** models). Although the hydrogen bond
 240 network in SBR becomes much denser than that in SBN, the calculation results show
 241 a smaller O...H binding energy difference (-24.39 kcal/mol in SBR and -30.10
 242 kcal/mol in SBN). These relatively smaller O...H binding energies (according to the
 243 total binding energies) imply that the major binding contributions are from O...Ag⁺
 244 interactions, further highlighting the critical role of open Ag⁺ sites in the selective
 245 removal of ReO₄⁻/TcO₄⁻.

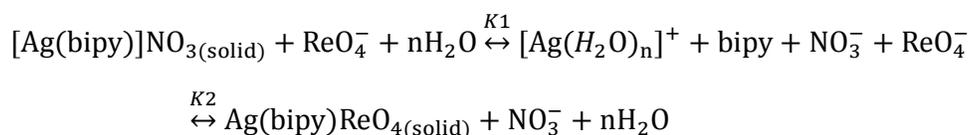


246

247 **Figure 3.** Fragment crystal structures of SBN (a) and SBR (c); A view of O...H bond and O...Ag⁺
 248 bond networks in SBN (b) and SBR (d); Atom colors: Ag = orange, N = blue, C = grey, H = white,
 249 Re = green, O = red, O...H bond = blue/white bar, O...Ag⁺ bond = orange/red bar.

250

251 We propose the large energetic difference could further give rise to clear
 252 deviation on the thermodynamic behavior between SBN and SBR materials,
 253 especially the solubility in aqueous solution, which can be utilized for the elucidation
 254 of the selective anion-exchange process described by the **Equation (1)**.⁴⁵ Where L
 255 denotes the ligand (4,4'-bipyridine), K_1 and K_2 represent the thermodynamic
 256 equilibrium constant.



257 (1)

258 The corresponding experimental standard thermodynamic equilibrium constant (K^θ),
259 reaction quotient (Q) and Gibbs free energy ($\Delta G(T)$) values are present in **Table S6**
260 and **Figure S10**. The thermodynamic driving force for ReO_4^- removal is attributed to
261 the much lower solubility product constant of SBR compared to that of SBN. For the
262 majority of pH conditions investigated (**Figures S10 and S11**), the SBR is less soluble
263 by at least three orders of magnitude, resulting in a spontaneous recrystallization
264 based phase transformation process. Furthermore as shown in **Table S7**, the solubility
265 product of SBR is as small as 2.16×10^{-13} , the smallest value among all known
266 perrhenate and pertechnetate salts up to date.

267 Based on the large thermodynamic gap, the anion-exchange reaction appeared
268 to be irreversible. Impressively as shown in **Figure S12**, the structure of SBR
269 remained identical even after immersed into a concentrated (1 M) NaNO_3 solutions
270 for 24 h while almost no noticeable amount of ReO_4^- leaks back to the aqueous
271 solution. With the help of heating to 60 °C, SBR was not able to turn back to SBN,
272 suggesting SBR is extremely robust in aqueous solution even containing huge excess
273 of NO_3^- . We also checked the stability of SBR materials in the presence of excess of
274 H_2PO_4^- and CO_3^{2-} to be excellent as demonstrated by the dissolution studies and
275 PXRD analysis on the solid sample after the dissolution studies (**Figure S13 and**
276 **Table S8**). In addition, SBR is thermally stable up to 380 °C (**Figure S14**), which is
277 again significant for technetium waste form design.

278

279 **Environmental Implications.** Searching for functional materials to efficiently
280 remove TcO_4^- and host lattices to trap technetium as waste forms are both highly
281 desirable and remains a significant challenge. Although different types of
282 precipitation agent have been used for separating and trapping ^{99}Tc , there is
283 significant opportunity for designing improved strategies and describing the

284 underlying molecular scale mechanisms that govern ^{99}Tc separation.⁴⁶⁻⁴⁸ More
285 importantly, previous investigation on the inorganic host lattice such as sodalite show
286 incorporation preference towards NO_3^- over $\text{ReO}_4^-/\text{TcO}_4^-$. This phenomenon is typical
287 following the well-known "Hofmeister bias"⁴⁹ because NO_3^- possesses higher charge
288 density and smaller size than $\text{ReO}_4^-/\text{TcO}_4^-$ and therefore higher affinity with the
289 inorganic host lattice. What we show here is a low-dimensional coordination polymer
290 materials equipped with abundant of open Ag^+ sites that are able to reverse the
291 "Hofmeister bias" and shows an excellent anion-exchange selectivity towards
292 $\text{ReO}_4^-/\text{TcO}_4^-$ over NO_3^- . This material is capable of both removing and immobilizing
293 TcO_4^- through a thermodynamically favorable phase transformation process, whose
294 molecular mechanism is clearly elucidated by single crystal structures before and after
295 uptake of the surrogate ReO_4^- , DFT calculation analysis on bonding, and solubility
296 tests. We believe this study may provide a new viewpoint on how to deal with ^{99}Tc
297 and SBR material should be further considered and evaluated for long-term storage of
298 technetium without the concern of the redox activity.

299

300 ■ ASSOCIATED CONTENT

301 Supporting Information

302 Details of materials and methods, additional data for SBN and SBR, are including
303 nine figures and six tables. The Supporting Information is available free of charge via
304 the Internet at <http://pubs.acs.org>.

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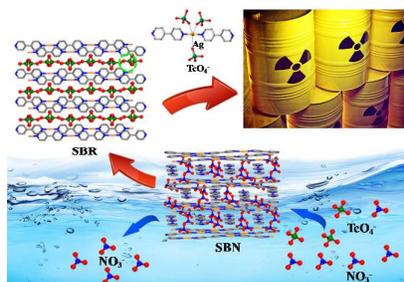
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476 **Exceptional Perrhenate/Per technetate Uptake and Subsequent Immobilization**
477 **by a Low-Dimensional Cationic Coordination Polymer: Overcoming the**
478 **Hofmeister Bias Selectivity**

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